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Spatial variation of mercury levels in nesting Bonelli's eagles from Southwest Portugal: effects of diet composition and prey contamination

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“Capsule”: *The effects of diet composition and prey contamination added up to determine the spatial variation of Hg levels in breeding Bonelli's eagles.*

Abstract

Mercury (Hg) was determined in adult Bonelli's eagles (*Hieraaetus fasciatus*) and their avian prey, from samples of feathers collected between 1992 and 2001 at the nesting sites of 21 pairs in Southwest Portugal. Eagle Hg levels showed great variation, reflecting primarily differences in diet composition and food chain biomagnification. Concentrations were positively correlated with the dietary proportion of insectivorous and omnivorous birds (e.g. egrets, corvids and thrushes), with very low levels for pairs feeding mainly on herbivores (e.g. rabbits, pigeons and partridges). Differences in prey contamination among breeding territories added to dietary effects in determining variation of Hg levels in eagles, shaping a spatial pattern that was largely consistent with a source of contamination in a coal-burning power-plant lying upwind of the study area. Despite this presumed contamination, Hg levels seemed to be of little concern to this eagle population, though there might be subtle deleterious effects on the reproductive output of a few pairs. This study emphasizes the need to account for dietary effects when biomonitoring Hg contamination using birds of prey.

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Keywords: Biomagnification; Biomonitoring; Birds of prey; Jay; Partridge; Pigeon

1. Introduction

As top predators, birds of prey are exposed to an array of persistent environmental contaminants that biomagnifies through food webs, especially organochlorine pesticides, polychlorinated biphenyls (PCBs), and mercury (Hg). Accumulation of these chemicals has been particularly well documented for aquatic food webs, where species such as sea eagles (*Haliaeetus* spp.) and ospreys (*Pandion haliaetus*) have shown poor breeding

and enhanced mortality in association with high pollutant burdens (Helander et al., 1982; Wiemeyer et al., 1984, 1988). Although much less documented, population declines attributed to environmental contaminations have also been shown for species feeding on terrestrial food chains such as the sparrowhawk (*Accipiter nisus*) (Newton et al., 1993). Because of this vulnerability to a variety of contaminants, birds of prey have been used extensively as biomonitors of environmental quality (Berg et al., 1966; Lindberg and Odsjö, 1983; DesGranges et al., 1998; Mañosa et al., 2003).

Besides their high trophic status, many birds of prey are territorial, non-migratory and long-lived, and so pollutant burdens recorded in body soft tissues, bones,

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¹ Luís Rocha Monteiro (1962–1999).

48 feathers and eggs are likely to reflect chemical contam-
 49 ination within their extended home ranges. This view
 50 underlies most biomonitoring programs, which assume,
 51 often implicitly, that spatial or temporal variations in
 52 pollutant burdens are coupled with comparable spatial or
 53 temporal trends in environmental contamination. Al-
 54 though this assumption may sometimes be warranted for
 55 birds of prey (e.g. DesGranges et al., 1998), there are at
 56 least some circumstances in which it may fall short of
 57 reality. A major source of potential shortcomings is
 58 related to diet composition, which may elicit variation in
 59 pollutant burdens among individuals of the same species
 60 collected at different locations or at different times,
 61 irrespective of corresponding variation in environmental
 62 contamination. For instance, some studies have linked
 63 local peak contamination levels in bald eagles (*Haliaeetus*
 64 *leucocephalus*), golden eagles (*Aquila chrysaetos*) and
 65 peregrine falcons (*Falco peregrinus*) with a high con-
 66 sumption of aquatic birds such as waders and seabirds
 67 (Lindberg and Odsjö, 1983; Parrish et al., 1983; Furness
 68 et al., 1989; Anthony et al., 1999). Similar effects for
 69 species feeding exclusively on terrestrial prey are scarce,
 70 though recent evidence suggests that they may also occur
 71 (Mañosa et al., 2003). Clearly, there is a need to evaluate
 72 in more detail the effects of diet composition on the
 73 pollutant burdens of birds of prey feeding on terrestrial
 74 food chains, and how these may confound the inter-
 75 pretation of spatial or temporal patterns in environ-
 76 mental contamination.

77 The present study addresses these issues, by analyzing
 78 the relationships between diet composition, prey con-
 79 tamination, and spatial variation of Hg levels in feathers
 80 of Bonelli's eagles (*Hieraetus fasciatus*). These are
 81 medium-sized eagles, whose numbers and range have
 82 declined markedly in Europe, where they are restricted to
 83 the Mediterranean region (Rocamora, 1994). Bonelli's
 84 eagles feed primarily on terrestrial birds and mammals,
 85 showing significant geographical variation in diet com-
 86 position depending on local habitat conditions (Real,
 87 1991). The study was carried out in the uplands of south-
 88 western Portugal, where a dense Bonelli's eagle popula-
 89 tion of great conservation significance lies downwind of
 90 a coal-burning power-plant. Because of this, there were
 91 concerns that these eagles could be exposed to an
 92 important source of Hg contamination, with potential
 93 negative repercussions upon their reproductive output
 94 and health condition. This justified a closer examination
 95 of factors underlying spatial variation in Hg burdens in
 96 the eagles and their prey.

97 2. Materials and methods

98 2.1. Study area

99 Data were collected as part of a long-term study on the
 100 Bonelli's eagle in the uplands of Algarve and western

Alentejo (southern Portugal), from 21 out of 25 eagle 101
 territories occupying about 3000 km² in a rough triangle 102
 linking the mountains of Cercal (341 m), Monchique 103
 (902 m) and Caldeirão (589 m) (Fig. 1). The hilly 104
 landscape is predominantly covered by cork oak 105
 (*Quercus suber*) woods, dense Mediterranean scrub and 106
 eucalyptus (*Eucalyptus globulus*) plantations, with sparse 107
 human occupation. Bonelli's eagles breed primarily in 108
 large cork oaks, eucalyptus and pine trees (*Pinus* spp.), 109
 and feed on domestic doves (*Columba labia*), red-legged 110
 partridges (*Alectoris rufa*), jays (*Garrulus glandarius*), 111
 rabbits (*Oryctolagus cuniculus*), and many other second- 112
 ary prey (Palma, 1994; L. Palma, unpublished data). The 113
 main potential source of Hg contamination is a coal- 114
 burning power-plant located at Sines, on the north- 115
 west corner of the study area (Freitas et al., 1999). No 116
 additional sources of Hg contamination, either telluric 117
 or agricultural, were identified within the study area. 118

2.2. Sampling procedures 119

2.2.1. Feather samples 120

121 From 1992 to 2001, shed feathers of adult Bonelli's
 122 eagles and feathers from avian prey remains were col-
 123 lected from nests and neighbouring tree perches to
 124 measure Hg levels. Active nests were visited three
 125 times during each breeding season, between the end of
 126 incubation and shortly after nest abandonment (March-
 127 July), and feathers of each species were collected in
 128 separate labelled plastic bags and stored in a freezer at
 129 -20 °C. Eagle feather samples were obtained on only
 130 2.3 ± 1.1 SD (1-4) years per breeding pair, because the
 131 location of some nests was unknown in early years of the
 132 study, some pairs did not breed every year and shed
 133 feathers were occasionally absent. Most eagle feathers
 134 were probably from females, as they tend to spend far
 135 more time near nests than males (Blondel et al., 1969;
 136 Morvan and Dobchies, 1987; L. Palma, unpublished
 137 data), and because the matching between sampling and
 138 moulting periods was closer for females than for males
 139 (L. Palma, unpublished data). Feathers were used as
 140 monitoring units because Hg in feathers reflects body Hg
 141 burden (Furness et al., 1986; Thompson et al., 1990) and
 142 it is almost entirely in the mono-methylated form
 143 (Thompson and Furness, 1989a,b). Furthermore, feath-
 144 ers have been widely used to monitor Hg levels in
 145 freshwater, marine and terrestrial bird species (Furness,
 146 1993), including birds of prey (Dauwe et al., 2003). Only
 147 body feathers were analysed, since they provide more
 148 representative samples for estimating whole-bird Hg
 149 content than flight feathers (Furness et al., 1986).

2.2.2. Diet composition 150

151 The diet of eagles in each individual breeding territory
 152 was analysed from prey remains collected during the
 153 visits to active nests and surrounding perches. Although

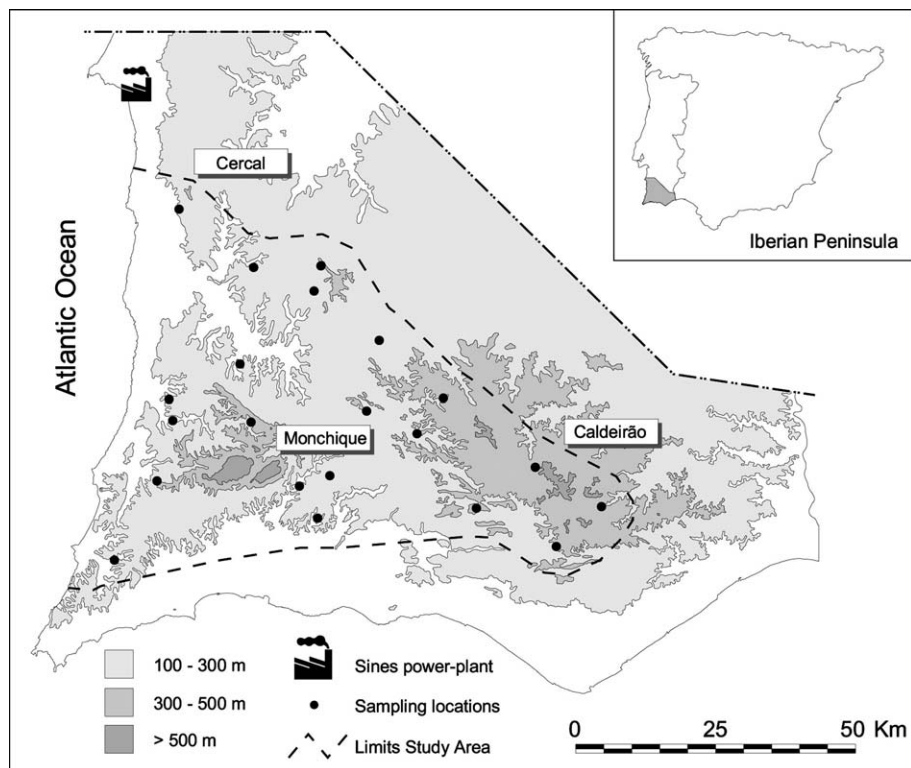


Fig. 1. Locations in Southwest Portugal where feather samples of Bonelli's eagles and their main avian prey were collected for the analysis of Hg contamination (1992–2001). Main mountain ranges are indicated.

154 remains correspond primarily to prey consumed by
 155 nestlings, they likely reflect also the diet of adults, which
 156 regularly eat part of the prey delivered to the nests
 157 (Blondel et al., 1969; Morvan and Dobchies, 1987;
 158 L. Palma, unpublished data). Remains were identified
 159 with the help of keys to bird feathers and a reference
 160 collection, and the minimum number of individuals in any
 161 sample was estimated from the highest number of
 162 identical bones of each prey type. This method tends to
 163 underestimate the consumption of small prey yielding few
 164 remains, while overestimating large prey or prey with
 165 a large proportion of rejected parts, such as bird feathers
 166 (Real, 1996). However, the method may be considered
 167 useful in comparative studies like this one, which aim to
 168 detect variation in the relative consumption of different
 169 prey, and not estimating the absolute diet composition.
 170 Because Hg contamination is strongly dependent on
 171 trophic position (Dietz et al., 2000), prey items were
 172 categorised according to whether they feed predomi-
 173 nantly on plants or animals. Species such as rabbits,
 174 partridges, wildfowl and seed-eating passerines were
 175 classified as primary consumers, whereas species such as
 176 egrets, gulls, birds of prey, corvids and other insectivore
 177 passerines were classified as secondary consumers.

178 2.2.3. Mercury determinations

179 Feather samples were analysed for total Hg concen-
 180 tration by Cold Vapour Atomic Absorption Spectros-

copy (CV-AAS). Samples were digested in a water bath 181
 at 70 °C for 6 h by the addition of concentrated H₂SO₄. 182
 After this period 5% KMnO₄ was added and the 183
 solution kept at 70 °C for two more hours. The KMnO₄ 184
 in excess was reduced with 20% NH₂OH.HCl. All 185
 reagents used throughout the work were of analytical 186
 grade. The glassware was previously decontaminated by 187
 immersion in an HNO₃ 1:5 solution and then washed 188
 with deionized water. Reproducibility was checked by 189
 performing successive measurements with the same 190
 sample. Relative standard deviations in the range 191
 3–5% were found. Accuracy of the method was within 192
 10% and was monitored analysing reference materials: 193
 tuna muscle 350 (International Atomic Energy Agency, 194
 Monaco) and RM50 (USA National Bureau of Stand- 195
 ards for Biological Material). Minimum detection 196
 limits (MDL) of 0.01 µg Hg/g digested sample were 197
 quantified using the Kaiser–Currie method (Gibbons 198
 and Coleman, 2001). Interferences due to matrix and the 199
 pre-treatment were assessed by the method of standard 200
 additions before the wet mineralization procedure. 201
 Recoveries of added Hg were close to 100%. Hg 202
 concentration is given on a wet weight basis. 203

204 2.2.4. Statistical analysis

205 Mean Hg concentrations were computed for feather
 206 samples collected from each Bonelli's eagle pair in any
 207 given year. The overall Hg concentration corresponding

to each breeding pair was then quantified as the mean of concentrations estimated in different years. Samples from avian prey were treated likewise. Before statistical analysis, nondetected measurements were replaced by half the detection limit (Gibbons and Coleman, 2001), and Hg concentrations were log-transformed to approach normality and homogenising variances (Zar, 1996). The arcsine transformation was used likewise for percentage data quantifying diet composition. Differences in Hg levels between species at matching locations were compared using paired-samples *t*-tests (Zar, 1996). Significance levels were corrected for multiple comparisons using the sequential Bonferroni technique (Rice, 1989). Pearson correlations and regression analyses were used to evaluate the relationships between eagle Hg levels, diet composition and avian prey contamination (Zar, 1996).

Spatial patterns in Hg levels for eagles and their main prey were mapped by interpolating to a continuous grid, the concentrations recorded at sampling locations, using inverse-distance weighing (Legendre and Legendre, 1998). Residuals of the regression equation relating eagle Hg levels to diet composition were also mapped, to illustrate the spatial contamination patterns after statistically accounting for dietary effects. In distance weighing, the extinction rule was $1/r^2$ (*r* is the distance between grid and sampling points), producing a smooth surface and avoiding the need to introduce an artificial cut-off distance (Legendre and Legendre, 1998).

3. Results

3.1. Diet

Eagle diets were described from an average 24.6 ± 15.2 SD (5–64) prey remains identified per eagle pair (Table 1). Almost half the overall remains were pigeons, over 95% of which were identified as domestic pigeons. Other important prey items were red-legged partridges, rabbits and corvids, about 75% of which were jays. Only 17.5% of prey remains corresponded to species categorised as secondary consumers, though their relative importance in the diet varied markedly among breeding pairs, from about 2.1% to 44.4%. Over 95% of individual prey identified corresponded to terrestrial species, with only gulls and mallards (*Anas platyrhynchos*) feeding regularly on aquatic food chains.

3.2. Eagle and avian prey Hg levels

Bonelli's eagles, red-legged partridges, domestic pigeons and jays showed some marked differences in their Hg concentrations (Table 2). Eagles showed much higher Hg concentrations than both partridges ($t_{18} = 7.513$, $P < 0.001$) and pigeons ($t_{19} = 9.822$,

Table 1

Composition of Bonelli's eagle diet in Southwest Portugal (1992–2001), as assessed from the remains of 541 identified preys recovered from the nests of 21 breeding pairs

Prey items		N	%
Birds			
Cattle egret	<i>Bubulcus ibis</i>	11	2.0
Gulls	<i>Larus</i> spp.	17	3.1
Red-legged partridge	<i>Alectoris rufa</i>	92	17.0
Domestic fowl	<i>Gallus gallus</i>	20	3.7
Pigeons	<i>Columba</i> spp.	256	47.3
Corvids	Corvidae	43	7.9
Thrushes	Turdidae	12	2.2
Other birds	Mainly Anatidae, Picidae and Strigidae	23	4.3
Mammals			
Rabbit	<i>Oryctolagus cuniculus</i>	67	12.4
Hare	<i>Lepus granatensis</i>	1	0.2

N = number of individual prey items; % = percentage of total prey recovered.

$P < 0.001$), but they had similar levels to those of jays ($t_{10} = 1.630$, $P > 0.1$). Likewise, levels in jays were much higher than in partridges ($t_9 = 7.434$, $P < 0.001$) and pigeons ($t_{10} = 8.195$, $P < 0.001$). Concentrations of Hg in partridges and pigeons were virtually identical ($t_{16} = 0.111$, $P > 0.9$).

3.3. Effects of diet and prey contamination

Relationships between eagle Hg levels and diet compositions were assessed using the 17 breeding pairs for which there were more than 10 prey remains. Hg concentrations were negatively correlated with the dietary proportion of pigeons ($r = -0.529$, $P < 0.05$), but not with those of partridges ($r = 0.233$, $P > 0.4$), jays ($r = 0.381$, $P > 0.1$) and rabbits ($r = 0.003$, $P > 0.9$). Analyses for other prey items were not made because they occurred too infrequently in eagle's diet. A strong positive correlation was found for prey categorised as secondary consumers ($r = 0.813$, $P < 0.001$), reflecting the strong influence of prey trophic position on eagle Hg levels (Fig. 2).

Concentrations of Hg in eagles were correlated with those in jays ($r = 0.634$, $P < 0.05$, $n = 11$), but not with

Table 2

Means, standard deviations and ranges of Hg concentrations ($\mu\text{g g}^{-1}$ wet weight) in feather samples of Bonelli's eagles and their main avian prey collected in Southwest Portugal (1992–2001)

Species	N	Mean	Standard deviation	Range
Bonelli's eagle	21	1.94	1.54	0.25–5.42
Domestic pigeon	20	0.13	0.17	<MDL–0.70
Red-legged partridge	18	0.11	0.11	<MDL–0.46
Jay	11	1.58	0.71	0.83–3.41

N = number of breeding pairs from which samples were collected; MDL = Minimum Detection Limit.

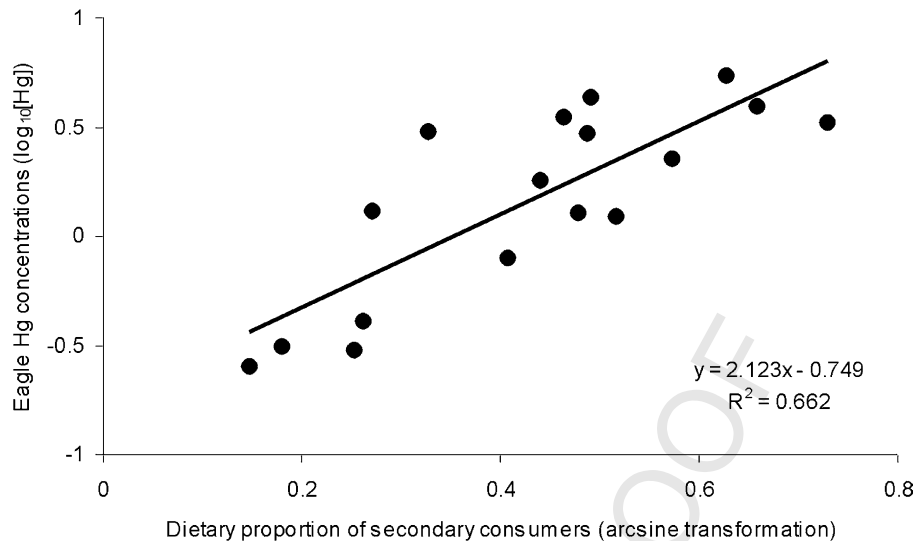


Fig. 2. Mean Hg concentrations (log-transformed; y) recorded in 17 Bonelli's eagle pairs breeding in Southwest Portugal (1992–2001), as a function of the dietary proportion (arcsine-transformed; x) of secondary consumers.

280 those in either partridges ($r = -0.064$, $P > 0.7$, $n = 19$)
 281 or pigeons ($r = 0.190$, $P > 0.4$, $n = 20$). In a multiple
 282 regression accounting for both the effects of diet
 283 composition and prey contamination, variation in Hg
 284 levels in eagles could be explained to a very large extent
 285 by the positive effects of the dietary proportion of
 286 secondary consumers and the concentration of Hg in
 287 jays (Table 3). Hg concentrations in pigeons and
 288 partridges never showed significant effects in similar
 289 multiple regressions relating eagle Hg levels with diet
 290 composition and prey contamination.

291 3.4. Spatial patterns

292 Hg levels in Bonelli's eagles tended to decline
 293 eastwards from the relatively high values recorded along
 294 the western portion of the coastal mountain ranges of
 295 Cercal and Monchique to the low values found in the
 296 eastern Caldeirão uplands (Fig. 3). However, high
 297 values were also found in one pair breeding along the
 298 north-eastern edge of the study area, and in two pairs
 299 breeding southeast of Monchique. The residuals of the
 300 regression equation between eagle Hg levels and the
 301 dietary proportion of secondary consumers (Fig. 2) were
 302 used to illustrate the spatial distribution of eagle
 303 contamination after correcting for dietary variation
 304 (Fig. 3). The emerging spatial pattern underlined the
 305 contrast between the western and eastern part of the
 306 study area, with the highest Hg levels concentrating
 307 around Monchique and the lowest in Caldeirão. High
 308 contamination values were also found in two pairs
 309 breeding on the north-eastern border of the study area.

310 For the three avian prey species there were differences
 311 in detail for the spatial patterns of Hg contamination,
 312 though they all showed a trend for higher values in the

western part of the study area (Fig. 3). Furthermore, the
 highest Hg levels in both pigeons and jays were recorded
 in the mountain of Cercal, in the sampling site closest to
 the industrial complex of Sines (Fig. 3). There were,
 however, exceptions to the west–east gradient of declining
 Hg levels, with some high values also recorded at the
 eastern end of Caldeirão for both pigeons and
 partridges.

4. Discussion

The Hg levels found in feathers of Bonelli's eagles
 breeding in the uplands of south-western Portugal
 showed great variation, which seemed to reflect primarily
 differences in diet composition and food chain
 biomagnification. The highest concentrations were
 recorded in pairs incorporating a high proportion of
 secondary consumers in their diet, whereas much lower
 values were found for eagles feeding almost exclusively
 on herbivores such as rabbits, pigeons and partridges.
 Comparable effects of trophic chain length have been

Table 3
 Multiple linear regression relating Hg concentrations in 10 Bonelli's eagle pairs breeding in Southwest Portugal, to prey contamination and diet composition ($R^2 = 0.954$, $F_{2,7} = 35.676$, $P < 0.001$)

Variables	Regression coefficients	t	P
Intercept	-0.678	-4.181	<0.01
Concentration of Hg in jays (log-transformed)	1.716	4.577	<0.01
Dietary proportion of secondary consumers (arcsine-transformed)	1.461	4.302	<0.01

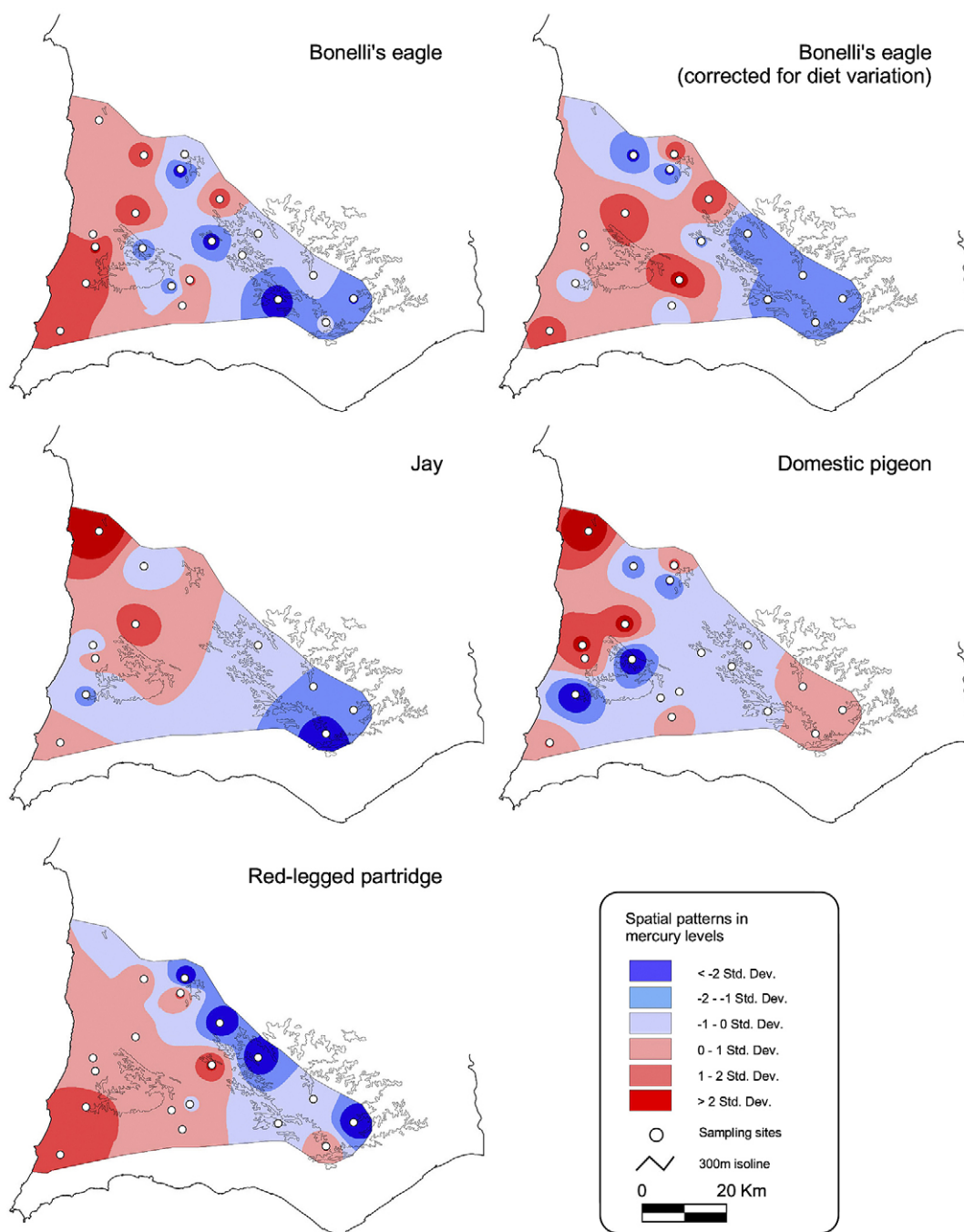


Fig. 3. Spatial distribution of Hg concentrations measured in Bonelli's eagles and their main avian prey. Values are given as standard deviations from the mean, to increase comparability among maps. Eagle data corrected for diet composition are the residuals of the linear regression depicted in Fig. 2, between eagle Hg levels and the dietary proportion of secondary consumers.

332 noted mainly in marine and freshwater systems (e.g.
 333 Elliot et al., 1996; Anthony et al., 1999), with
 334 comparable data generally lacking for terrestrial food
 335 webs. In a study involving organochlorine contaminants
 336 in goshawk (*Accipiter gentilis*) eggs, however, Mañosa
 337 et al. (2003) also documented the highest concentrations
 338 in association with a higher consumption of passerine
 339 birds relative to that of rabbits. The paucity of data for
 340 terrestrial chains is probably related to their shorter

length in relation to that of aquatic ones, which lessens
 the potential for Hg biomagnification along the food
 web (Dietz et al., 2000). Nevertheless, this study strongly
 suggests that food web length in terrestrial systems may
 also be a major source of variation in Hg contamination
 for top predators such as the Bonelli's eagle, which can
 feed at multiple trophic levels.

After statistically accounting for dietary effects, Hg
 concentrations in eagles also reflected the contamination

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level in some of their prey. Indeed, there was a strong correlation between Hg concentrations in feathers of eagles and jays, though no such relationship was apparent for pigeons and partridges. The general pattern for other secondary consumers might have been similar to that recorded for jays, though the shortage of feather samples precluded the testing of this hypothesis. Lack of relationship between eagle and pigeon contamination was unexpected, as these are the staple food of eagles. However, Hg levels were very low in herbivorous prey, which suggests that eagles acquire most of their burden through the intake of secondary consumers. Additional information on the sources of variability in prey contamination is needed to gain a better understanding of the mechanisms leading to Hg accumulation in the eagles.

The strong relationship between eagle Hg levels, diet composition and contamination of prey collected from nests, suggests that concentrations found in shed body feathers probably resulted primarily from exposure during the breeding season. In the study area, adult eagles are largely resident within the breeding territories, starting to visit the nests in November, long before the shed feathers could be found, and remaining in the surroundings at least until juvenile emancipation in August–September (L. Palma, unpublished data). The shed body feathers analysed were generally collected in the late nestling and early fledging periods (>85% in April–June), corresponding to the post-nuptial moult, which may extend until early autumn (Parellada, 1984; L. Palma, unpublished data). These shed feathers grew during the previous moulting season, thus receiving Hg that had been stored in body tissues over the preceding months (Furness et al., 1986; Furness, 1993; Dauwe et al., 2003). Hg probably accumulated in the adult eagles mainly while foraging within their extended breeding ranges, thus integrating contamination from areas lying in general within 10 km from the nests (L. Palma, unpublished data). This supports the assumption that variation among pairs in the concentrations recorded in shed feathers should reflect at least partly the broad scale spatial trends in environmental contamination, once the dietary effects are accounted for.

Regional Hg concentration trends in prey species and in Bonelli's eagles, after correcting for dietary variation, broadly agree with the hypothesis of a contamination source in the industrial complex of Sines, presumably associated with the coal-burning power-plant, though there were differences among species. Reasons for these differences are unclear, but they may probably be attributed to local factors and sampling variability, which likely added to the large scale contamination trend in influencing the spatial patterns observed. These local factors are impossible to assess with the data collected, but they may result from variation among eagle breeding territories concerning the diets or feeding habitats of the

prey species captured. Despite these confounding factors, there was a general trend for higher Hg levels in the western uplands of Cercal and Monchique, which lie immediately downwind of the industrial complex and are thus probably more likely to be contaminated from airborne pollutants than the eastern Caldeirão mountains. Furthermore, precipitation along the coastal uplands, particularly in Monchique, is in general much higher than further inland, which may favour the removal from the atmosphere and local wet deposition of Hg emitted in combustion facilities (Carpi, 1997). This view is also supported by the distribution in lichens of pollutants presumably originating from the Sines coal-powered electric plant, namely Hg, sulphur and selenium, which tended to show higher concentrations in the western uplands than in the east (Freitas et al., 1999). Although comparable patterns were not readily apparent in a similar study using mosses (Figueira et al., 2002), these results call for a more detailed examination of the distribution and biological effects of contaminants emitted from Sines up to several tens of kilometres from the source. This is particularly important in the case of Hg, which biomagnify through food chains and may negatively affect endangered top predators such as the Bonelli's eagle.

Although the mean Hg contamination recorded in eagles can be considered generally low, the highest levels detected might be of concern regarding eventual adverse impacts on the breeding productivity of some individual pairs (Berg et al., 1966; Lindberg and Odsjö, 1983; Parrish et al., 1983; Movalli, 2000). Establishing a benchmark for critical Hg concentrations in feathers is difficult, however, because Hg bonded to keratin and sequestered in feathers no longer represents a risk to the bird (Furness, 1993), and its levels may be uncorrelated with concentrations in eggs (e.g. DesGranges et al., 1998). Nevertheless, Hg concentrations in eagle feathers reported in this study, were correlated with those found in a small sample of added eggs ($n = 13$) collected from 10 breeding territories in a concurrent study (Blanco, 2001). There was a strong linear relationship between Hg levels in feathers and eggs from individual pairs ($R^2 = 0.772$, $F_{1,8} = 27.078$, $P < 0.001$), with feather levels of $4.1 \mu\text{g g}^{-1}$ corresponding to eggs containing the benchmark of $1.0 \mu\text{g g}^{-1}$ (wet weight). This concentration may be the lowest associated with deformities of particularly sensitive embryos, though it is unlikely to affect more than a small percentage of eggs (Heinz and Hoffman, 2003). In this study, only two out of 21 Bonelli's eagle pairs (9.5%) showed feather levels in excess of this threshold ($4.3\text{--}5.4 \mu\text{g g}^{-1}$), and may thus be considered moderately susceptible to reproduction impairment due to Hg contamination. For the overall breeding population, however, it is unlikely that Hg contamination can negatively affect the reproductive output.

462 Results from this study add to a body of evidence
463 derived primarily from aquatic food webs, suggesting
464 that diet variation may have major confounding effects
465 in studies biomonitoring environmental contamination
466 using birds of prey (Anthony et al., 1999; Mañosa et al.,
467 2003). To overcome potential shortcomings, some
468 authors recommended that bird species with narrow
469 and inflexible diets should be used in contamination
470 studies, rather than generalist feeders (Monteiro and
471 Furness, 1995). However, true dietary specialists are
472 probably hard to find, and so the critical assumption of
473 constant diets across space and time may frequently be
474 unwarranted. A detailed knowledge of diet variation
475 and the statistical control of dietary influences, as in this
476 study, may thus be generally required to derive
477 meaningful trends in Hg environmental contamination
478 from the corresponding spatial or temporal variation in
479 concentrations recorded in birds of prey.

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